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The structure of the bacterial and archaeal community in a biogas digester as revealed by denaturing gradient gel electrophoresis and 16S rDNA sequencing analysis

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Keywords

16S rDNA, anaerobic digestion, biogas digester, denaturing gradient gel electrophoresis (DGGE), microbial community, pig manure.

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Abstract

Aims: To identify the bacterial and archaeal composition in a mesophilic biogas digester treating pig manure and to compare the consistency of two 16S rDNA-based methods to investigate the microbial structure.

Methods and results: Sixty-nine bacterial operational taxonomic units (OTU) and 25 archaeal OTU were identified by sequencing two 16S rDNA clone libraries. Most bacterial OTU were identified as phyla of *Firmicutes* (47·2% of total clones), *Bacteroides* (35·4%) and *Spirochaetes* (13·2%). *Methanoculleus bourgensis* (29·0%), *Methanosarcina barkeri* (27·4%) and *Methanospirillum hungatei* (10·8%) were the dominant methanogens. Only 9% of bacterial and 20% of archaeal OTU matched cultured isolates at a similarity index of ≥97%. About 78% of the dominant bacterial (with abundance >3%) and 83% of archaeal OTU were recovered from the denaturing gradient gel electrophoresis (DGGE) bands of V3 regions in 16S rDNAs.

Conclusions: In the digester, most bacterial and archaeal species were uncultured; bacteria belonging to *Firmicutes*, *Bacteroides* and *Spirochaetes* seem to take charge of cellulolysis, proteolysis, acidogenesis, sulfur-reducing and homoacetogenesis; the most methanogens were typical hydrogenotrophic or hydrogenotrophic/aceticlastic; DGGE profiles reflected the dominant microbiota.

Significance and Impact of the Study: This study gave a first insight of the overall microbial structure in a rural biogas digester and also indicated DGGE was useful in displaying its dominant microbiota.

Introduction

As the biggest developing country in the world, China has a population of 1·3 billion. Among these people, 72·3% of them are living in the rural areas. Because of the rapid growth of industry and energy consumption in the city, energy shortage becomes a big problem [The Priority Programme for China's Agenda 21; see the World Wide Web (WWW) site: http://www.acca21.org.cn/indexe8. html]. Furthermore, the demands of energy in the rural areas also increase daily. On the other hand, there is a great deal of waste biomass produced in the countryside, including a large amount of lignocellulose byproducts in agriculture and excrement of poultry and cattle, which

creates problems for the rural environment. To improve the environmental and living conditions in the countryside and to create a sustainable development in rural economy (Gupta 2003), it is necessary to resolve the energy and sanitary problems in the rural areas.

Anaerobic fermentation of waste biomass not only generates biogas fuel for cooking, lighting and heating, but also reduces waste biomass, potentially providing a mutually beneficial situation for the environmental, social and cycling-economic development in the rural areas. The technology for constructing biogas digesters at different scales and for different applications for treating rural wastes is well established (Chynoweth *et al.* 1999), while the efficiency of biogas production needs to be improved

urgently. The composition of the microbial community in a biogas digester directly determines its efficiency and biogas yield. The process of anaerobic conversion of different biomasses to methane usually includes four steps: hydrolysis, acidogensis, acetogenesis and methanogenesis, in which hydrolytic, fermentative bacteria, acetogens and methanogens play distinct roles, respectively (Pretty *et al.* 2002; Angenent *et al.* 2004).

Hydrolytic and acidogenic bacterial strains identified as Clostridium thermocellum, Clostridium leptum, Clostridium botulinum, Bacteroides termitidis, Desulfovibrio desulfuricans, Treponema palladium and Pirochaeta aurantia, acetogens such as Syntrophobacter wolinii and Syntrophomonas wolfei and archaeal methanogens belonging to Methanosaeta sp., Methanocorpusculum sp., Methanoculeus sp. and Methanobrevibacter sp. Methanobacterium sp., Methanosarcina sp. and Methanobrevibacter sp. were isolated from biogas digesters and anaerobic storages of animal manure in several studies (Boone and Bryant 1980; Zhao et al. 1986; Ney et al. 1990; Meher and Ranade 1993; Ohkuma and Kudo 1996; Whitehead and Cotta 1999; Cotta et al. 2003; Snell-Castro et al. 2005; Drake et al. 2006). A number of studies have indicated that only a small portion of the microorganisms (0·1–25%) were cultured (Cotta et al. 2003).

Culture-independent approaches, mainly 16S rDNAbased methods, e.g. cloning and sequencing, D/TGGE (denaturing/temperature gradient gel electrophoresis) fingerprinting, FISH (fluorescence in situ hybridization), RFLP [restriction fragment length polymorphism, or ARDRA (amplified rDNA restriction analysis)] and T-RFLP (terminal-RFLP), are applied to elucidate the diversity and composition of microbial communities in anaerobic methanogenic digesters widely used for the treatment of municipal and industrial wastewater (Karakashev et al. 2005; Connaughton et al. 2006; Mladenovska et al. 2006; Cirne et al. 2007; Klocke et al. 2007; Sousa et al. 2007a; Ye et al. 2007). Connaughton et al. (2006) and Sousa et al. (2007a) reported changes over time of the bacterial and archeal population in an anaerobic digester, the former compared the biomass composition with the activity of the digester and the latter compared the composition of microbiota in the presence and absence of long chain fatty acids. Ye et al. (2007) reported changes in digesters operated at several pH. Several studies involved investigation of microbial compositions in anaerobic methanogenic digesters treating rural waste (Mladenovska et al. 2006; Cirne et al. 2007; Klocke et al. 2007). By TGGE and cloning library and sequencing, Mladenovska et al. (2003) found that the most dominant methanogens in lab-scale anaerobic digesters with cattle manure or a mixture of cattle manure with glycerol trioleate were phylogenetically related to Methanosarcina siciliae. The bacterial and archaeal composition identified

by T-RFLP analysis of 16S rRNA genes were found to be identical in two thermophilic continuously stirred tank reactors (CSTR), treating nontreated manure and pretreated manure for 40 min at 140°C, respectively (Mladenovska *et al.* 2006). However, to our knowledge, no previous studies characterized the overall microbial communities in a biogas digester treating rural waste.

Compared with the sequencing analysis of 16S rDNA genes in clone libraries, DGGE profiles of the V3 region was simpler and less costly for analysing the structural variation between different microbial systems or the spatio-temporal dynamics of the same system (Muyzer et al. 1993). It could also be useful in tracing the variation of dominant microbial organism in a biogas digester. However, previous studies have indicated that DGGE profiles were not consistent with the results from the sequencing of clone libraries in different microbial systems (Krave et al. 2002; Freeman et al. 2008; Wakase et al. 2008). It is important to ensure whether the results from DGGE profiles are consistent with clone library analyses before they are applied in biogas digesters.

For the purpose of both basic research and biogas biotechnology, there is considerable interest in elucidating the microbial composition and metabolic diversity involved in biogas production, as well as setting up an applied and less costly method to trace the variation of microbial structure in anaerobic bio-digesters. In this study, the overall microbial communities in a mesophilic anaerobic biogas digester were investigated by analysing the diversity of 16S rDNA using DGGE and sequence analysis. Such an investigation for the composition of the microbial community in the biogas digesters would be the first step to elucidate the relationship between the efficiency of biogas production in the digesters and the structure and variation of the microbiota. The purpose of this study was twofold: (i) to characterize the microbial diversity of an anaerobic biogas digester using cultureindependent methods and (ii) to compare the discriminatory power of DGGE vs 16S library screening.

Materials and methods

Sample collection

A biogas slurry sample (3 l) was collected in June 2005 from a biogas digester in Qianwei Village, Chongming County, Shanghai, China, which was built in the early of 1980s and kept running for about 25 years. The digester has a volume of 600 m³ for anaerobic fermentation and produces 150 m³ biogas per day. The main fermentation substrate is pig manure from a nearby piggery breeding 1500 pigs. The digester can treat 1800 tonnes of pig manure from the piggery every year (Hao and Liu 2006).

DNA extraction

To remove extracellular DNA and soluble organic contaminants, the fresh slurry samples (50 ml) were washed thrice in five volumes of TENP buffer [50 mmol l⁻¹ Tris-HCl, 20 mmol l⁻¹ EDTA, 100 mmol l⁻¹ NaCl, 0·01 g ml⁻¹ polyvinylpyrrolidone (PVP), pH 10], twice in five volumes of sterile phosphate-buffered saline (PBS buffer, 137 mmol l⁻¹ NaCl, 2·7 mmol l⁻¹ KCl, 1·5 mmol l⁻¹ KH₂PO₄, 8·1 mmol l⁻¹ Na₂HPO₄ in distilled water, pH 7·4) and vortexed for 5 min and centrifuged for 10 min at 10 000g. The washed cell pellets were resuspended in a suitable volume of sterile PBS buffer containing glycerol at a final concentration of 20%, divided into 4-ml aliquots, and stored at -70°C until nucleic acid extraction.

After thawing and centrifugation of subsamples (1 ml), the supernatants were removed. The genomic DNA was then extracted using QIAamp DNA stool mini kit (Qiagen, Heidelberg, Germany) according to the manufacturer's instruction with minor modification. The treated sample was carefully disrupted by Cell Disruptor Genie (Scientific Industries Inc., New York) in the 2·0-ml microcentrifuge tube after adding the extraction buffer. The purified DNA was quantified with a Biophotometer (Eppendorf), and stored at −20°C until use.

Establishment of bacterial and archaeal 16S rDNA libraries

Bacterial and archaeal clone libraries were generated from polymerase chain reaction (PCR)-amplified 16S rDNA using bacterial primers 27f (5'-GAG AGT TTG ATC CTG GCT CAG-3') and 1495r (5'-CTA CGG CTA CCT TGT TAC GA-3') (Bianciotto et al. 1996) and the archaeal primers 1Af (5'-TCY GKT TGA TCC YGS CRG AG-3') and 1100Ar (5'-TGG GTC TCG CTC GTT G-3') (Embley et al. 1992). Reaction mixtures (25 μl) contained 2.5 μl of 10 × PCR buffer (TaKaRa Inc., Dalian, China), 0.2 mmol l⁻¹ each of deoxyribonucleotide triphosphates (dNTP), $0.5 \mu \text{mol l}^{-1}$ of each primer, 1 U of Ex Tag DNA polymerase (TaKaRa Inc.) and 5 ng of template DNA. PCR reactions were performed on a Flexigene thermal cycler (Techne Flexigene, Cambridge, UK). The PCR reaction for archaeal 16S rDNA was performed using the following programme: initial denaturation for 3 min at 94°C; 30 cycles of denaturation (1 min at 94°C), annealing (1 min at 55°C) and extension (2 min at 72°C) with a final extension of 72°C for 10 min. The optimized PCR amplification conditions for bacterial 16S rDNA were as follows: initial denaturation at 95°C for 1.5 min; 5 cycles of 95°C for 30 s, 60°C for 30 s, 72°C for 1.5 min; 5 cycles of 95°C for 30 s, 55°C for 30 s, 72°C for 1.5 min; 15 cycles of 95°C for 30 s, 50°C for 30 s, 72°C for 1·5 min and a final extension of 72°C for 10 min.

To minimize PCR artefacts, 'reconditioning PCR' was performed as described by Thompson *et al.* (2002) after the initial amplification of the bacterial and archaeal 16S rDNA as described before. The initial PCR-amplified reaction was diluted 10-fold in a fresh reaction mixture of the same composition and cycled thrice using this programme. ssDNA and heteroduplex DNA could be minimized by adding excess primer during the 'reconditioning PCR' (Zhang *et al.* 2005).

Cloning and sequencing

Exactly 100 ul of bacterial and archaeal 16S rDNA reconditioning PCR products were electrophoresed on 1.0% agarose and the bands of the correct size (c. >1.5 kb for bacteria, and 1.1 kb for archaea) were purified using 3S PCR Product Purification Kit V2:0 (Shenergy Biocolor Biological Science & Technology Co., China), respectively. Finally, the purified product was cloned into the pMD18-T plasmid vector (TaKaRa Inc.) following the manufacturer's instructions. The ligated products were transformed into Escherichia coli TOP10 competent cell (Invitrogen) with ampicillin and blue/white screening, and positive clones were arrayed in 96-well plates and stored at -80°C for long-term storage. Plasmid inserts were checked by PCR amplification using the M13 PCR set. Exactly 310 bacterial and 192 archaeal positive insertcontaining clones were randomly selected for gene sequencing. The template DNA was prepared from overnight cultures of selected clones using an alkaline miniprep kit (Qiagen), sequencing were performed on an ABI 3730 DNA sequencer with Big Dye terminator chemistry as specified by the manufacturer (Applied Biosystems).

16S rDNA V3 region amplification

40 mmol l^{-1} acetate, 1·0 mmol l^{-1} EDTA) and ethidium bromide (0·5 μ g ml⁻¹) under ultraviolet (UV) light.

Denaturing gradient gel electrophoresis

DGGE of the PCR products was performed by the method described by Muyzer et al. (1993) with the DCode Universal Mutation Detection system (Bio-Rad Laboratories, Hercules, CA < USA). Denaturing gradient gel [1 mm thickness \times 160 \times 160 mm; 1 \times TAE (40 mmol l⁻¹ Tris base with 1.0 mmol l-1 EDTA and 20 mmol l-1 sodium acetate at pH 7·4), 8% acrylamide-bisacrylamide (37·5:1), and 25-60% (35-70% for archaea, based on our unpublished results of perpendicular DGGE) denaturant $(100\% = 7 \text{ mol l}^{-1} \text{ urea with } 40\% \text{ formamide})]$ was poured with a gradient delivery system (model 475; Bio-Rad Laboratories). Electrophoresis was performed at a constant temperature of 60°C, first for 10 min at 25 V and then for 5 h at 200 V in 1 × TAE buffer. After the electrophoresis, the gels were stained with AgNO₃ as described by the manufacturer.

Clone library construction of single DGGE bands

Each gel slice that contained an obvious DNA band was excised with a clean razor blade and placed in an 1·5-ml Eppendorf tube. The gel slice was crushed and incubated with 50 μ l of TE buffer at 4°C overnight. The 3- μ l supernatant was subjected to a second PCR under the same conditions as described before. The re-amplified PCR products were examined by DGGE to confirm that single bands were present at the same positions. The PCR products were then purified with Mini-DNA Rapid Purification Kit (BioDev, Beijing, China) and cloned into the pMD18-T Vector (TaKaRa Inc.) to construct the clone libraries. Five clones were picked from each library and sent to Invitrogen (Shanghai, China) for sequencing.

Sequence analysis

Sequences were edited manually to remove vector and ambiguous sequences at the ends by scanning of the individual chromatograms using Chromas software ver.2.23 (Technelysium, Shanghai, China). Chimeras were checked by the CHIMERA_CHECK programme (Cole et al. 2003) in Ribosomal Database Project (RDP) at first, and then were further firmed by Bellerophon programme on the Greengenes website (DeSantis et al. 2006). All reference sequences were obtained from the GenBank and RDP. Then all the sequences and their closest relatives were fitted into an alignment using the automated tools of the ARB software package (Ludwig et al. 2004). Aligned sequences were added to the ARB neighbour-joining tree

(based on pairwise distances with Olsen correction) with the parsimony insertion tool as described by Ley et al. (2005). Sequences with internal regions of poor quality leading to alignment problems were excluded from further analysis. Dotur (Schloss and Handelsman 2005) was used to cluster sequences into operational taxonomic units (out) by % pairwise identity (%ID, using a furthest-neighbour algorithm and a precision of 0·03). The stability of tree branches was assessed by the bootstrap method using 1000 replicates.

Nucleotide sequence accession numbers

Bacterial nucleotide sequences obtained in this study are available in the GenBank database under accession numbers: EU358617–EU358650 and EU358676–EU358744. Archaeal nucleotide sequences obtained in this study are available in the GenBank database under accession numbers: EU358606–EU358616 and EU358651–EU358675.

Results

Microbial community analysed by clone library-bacterial community

The 16S rDNA genes were amplified from the total DNA extracted from the biogas slurry sample with a bacterial primer set 27f/1495r and amplicons were ligated to pMD-18 T vector to construct a library. In total, 310 clones were randomly selected and sequenced. After eliminating low-quality (68 clones) sequences and chimeric sequences (53 clones), 189 sequences were used for the following analyses. The coverage of the library was 81·0%, indicating that the library was large enough for further analyses. The 189 sequences of 16S rDNA genes were classified into 69 OTU (Table 1). The abundances of all OTU in the library were less than 7%, where only nine OTU were more than 3%. The accession number, sources and described functions of their phylogenetically closest matched organisms are also listed in Table 1.

Only 14 of the 69 OTU in the bacterial library were matched to the closest related known sequences deposited in NCBI and RDP at a similarity index of more than 97%, which was regarded as an experiential index for differentiating species (Table 1). Twenty-nine of the 69 OTU were matched to the closest related known sequences deposited in the databases at a similarity index of between 90% and 96%, while 26 of the 69 OTU were matched to the closest related known sequences at a similarity index of between 80% and 90%. Nearly 80% of the bacteria in this digester may be new, previously undescribed species. Even among the 14 OTU most closely related to known sequences, only 6 OTU matched with the cultured

Table 1 Taxonomic relationship of bacterial 16S rDNA sequences from Chongming biogas digester compared (BLAST) with public databases (RDP, Greengenes and NCBI)

OTU (%)	% of total	Accession number	Phylogenetically most closely related organism (accession no.)	Sm (%)	Phylum	Functional group	Source
BS01	6.9	EU358676	Uncultured anaerobic bacterium (AY953213)	87	Firmicutes	_	USA: swine lagoon
BS02	6.3	EU358677	Alkaliflexus imshenetskii (AJ784993)	90	Bacteroidetes	Acidogenic	Russia: soda lake
BS03	6.3	EU358678	Petrimonas sulfuriphila (AY570690)	90	Bacteroidetes	Acidogenic	Canada: biodegraded oil
BS04	5.3	EU358679	Proteiniphilum acetatigenes (AY742226)	94	Bacteroidetes	Proteolytic	China: sludge of UASB reactor
BS05	4.8		Spirochaeta sp. SPN1 (AJ698092)	88	Spirochaetes	_	Germany: hindgut of the termite
BS06	4.2	EU358681	Uncultured Clostridiaceae (DQ069192)	94	Firmicutes	_	USA: SA Au mine
BS07	3.7	EU358682	Clostridium quinii (X76745)	99	Firmicutes	Acidogenic	UK: type strain DSM6736
BS08	3.2	EU358683	Uncultured bacterium (EF559197)	99	Bacteroidetes	_	France: mesophilic digester
BS09	3.2	EU358684	Ruminofilibacter xylanolyticum (DQ141183)	91	Bacteroidetes	_	China: rumen
BS10	2.6	EU358685	Clostridium thermocellum (L09173)	86	Firmicutes	Cellulolytic	DSM 1237
BS11	2.6	EU358686	Uncultured <i>Clostridium</i> sp. (DQ309375)	93	Firmicutes	_	India: effluent treatment
BS12	2.1	EU358687	Treponema brennaborense (Y16568)	91	Spirochaetes	Acidogenic	FRG: dairy cow
BS13	2.1	EU358688	Treponema brennaborense (Y16568)	91	Spirochaetes	Acidogenic	FRG: dairy cow
BS14	2.1	EU358689	uncultured Fibrobacteres (EF454806)	90	Fibrobacteres	_	USA: termite hindgut
BS15	2.1	EU358690	Clostridium thermocellum (L09173)	88	Firmicutes	Cellulolytic	DSM 1237
BS16	2.1	EU358691	Paludibacter propionicigenes (AB078842)	88	Bacteroidetes	Acidogenic	Japan: rice straw in paddy soil
BS17	2.1	EU358692	Uncultured Bacteroidetes(EF111167)	89	Bacteroidetes	_	Colombia: bogota river
BS18	1.6	EU358693	Uncultured spirochete (EF562545)	94	Spirochaetes	_	Canada: biodegraded oil
BS19	1.6	EU358694	Clostridium orbiscindens (Y18187)	89	Firmicutes	Acidogenic	DSM 6740
BS20	1.6	EU358695	Clostridium bartlettii (AY438672)	92	Firmicutes	Acidogenic	USA: human feces
BS21	1.6	EU358696	Anaerovorax odorimutans (AJ251215)	92	Firmicutes	Acidogenic	Germany: strain NorPut
BS22	1.6	EU358697	Paludibacter propionicigenes(AB078842)	88	Bacteroidetes	Acidogenic	Japan: rice straw in paddy soil
BS23	1.6	EU358698	uncultured Cytophaga sp. (EF562564)	94	Bacteroidetes	_	USA: paper pulp column
BS24	1.1	EU358699	uncultured Verrucomicrobia (AM040118)	86	Verrucomicrobia	_	Germany: sandy sediments
BS25	1.1	EU358700	Tissierella praeacuta (X77848)	96	Firmicutes	_	DSM 5675
BS26	1.1	EU358701	Clostridium orbiscindens (Y18187)	87	Firmicutes	Acidogenic	DSM 6740
BS27	1.1	EU358702	Uncultured Clostridiales (AB234509)	89	Firmicutes	_	Japan: gut of termites
BS28	1.1	EU358703	Clostridium chartatabidum (X71850)	99	Firmicutes	Cellulolytic	DSM 5482
BS29	1.1	EU358704	Tissierella praeacuta (X80833)	93	Firmicutes	_	UK: type strain ATCC 25539
BS30	1.1	EU358705	Aminobacterium colombiense (AF069287)	86	Firmicutes	_	Australia: anaerobic sludge
BS31	1.1	EU358706	Uncultured Leptospiraceae (EF454914)	88	Spirochaetes	_	USA: termite hindgut
BS32	0.5	EU358707	Uncultured spirochete (EF562545)	94	Spirochaetes	_	Canada: biodegraded oil
BS33	0.5	EU358708	Sphaerochaeta sp. RCcp2 (DQ833401)	89	Spirochaetes	_	USA:TCE-dechlorinating
BS34	0.5	EU358709	Spirochaeta sp. SPN1 (AJ698092)	88	Spirochaetes	_	Germany: hindgut of the termite
BS35	0.5	EU358710	Xanthomonas vasicola (Y10755)	95	Proteobacteria	_	FRG: Strain LMG 736 T
BS36	0.5	EU358711	Uncultured planctomycete (DQ206406)	98	Planctomycetes	_	USA: soda lake water
BS37	0.5	EU358712	Anaerovorax odorimutans (AJ251215)	92	Firmicutes	Acidogenic	Germany: strain NorPut
BS38	0.5	EU358713	Clostridium straminisolvens (AB125279)	88	Firmicutes	Cellulolytic	Japan:cellulose-degrading
BS39	0.5	EU358714	Clostridiaceae bacterium 80Wc (AB078860)	95	Firmicutes	_	Japan: rice straw in paddy soil
BS40	0.5	EU358715	Uncultured Clostridiales (AB234479)	95	Firmicutes	_	Japan: gut of termites
BS41	0.5	EU358716	Sporobacter termitidis (Z49863)	91	Firmicutes	Homoacetogenic	Australia: wood-feeding termite
BS42	0.5	EU358717	Desulfotomaculum guttoideum (Y11568)	93	Firmicutes	Acetogenic	DSM 4024
BS43	0.5	EU358718	Garciella nitratireducens (AY176772)	89	Firmicutes	Acidogenic	Mexico: oilfield separator
BS44	0.5	EU358719	Streptococcus alactolyticus (AF201899)	99	Firmicutes	Acidogenic	Denmark: ATCC 43077
BS45	0.5	EU358720	Uncultured bacterium (AY976000)	98	Firmicutes	_	USA: human colon mucosal
BS46	0.5		Clostridium nexile (X73443)	93	Firmicutes	_	DSM 1787
BS47		EU358722	Tissierella praeacuta (X80833)	93	Firmicutes	_	UK: type strain ATCC 25539
BS48			Clostridium sp. (X75909)	98	Firmicutes	_	UK: strain BN II
BS49			Sporobacter termitidis (Z49863)	92	Firmicutes	Homoacetogenic	Australia: wood-feeding termite
BS50			Lactobacillus reuteri F275(CP000705)	99	Firmicutes	Acidogenic	DSM 20016

Table 1 (Continued)

OTU (%)	% of total	Accession number	Phylogenetically most closely related organism (accession no.)	Sm (%)	Phylum	Functional group	Source
BS51	0.5	EU358726	Leuconostoc citreum(AF111949)	100	Firmicutes	Acidogenic	South Korea: fermented cabbage
BS52	0.5	EU358727	Guggenheimella bovis (AY272039)	89	Firmicutes	Proteolytic	USA: bovine dermatitis digitalis
BS53	0.5	EU358728	Desulfotomaculum thermocisternum (U33455)	83	Firmicutes	Acetogenic	Norway: hot North Sea oil
BS54	0.5	EU358729	Uncultured bacterium (EF559146)	97	Firmicutes	_	France: mesophilic digester
BS55	0.5	EU358730	Gracilibacter thermotolerans (DQ117465)	88	Firmicutes	Acidogenic	USA: acid sulfate wetland
BS56	0.5	EU358731	Moorella glycerini (U82327)	81	Firmicutes	Homoacetogenic	USA: strain YS6
BS57	0.5	EU358732	Uncultured Thermoanaerobacteriales (AY684076)	92	Firmicutes	_	Germany: methanogenic enrichment
BS58	0.5	EU358733	Uncultured bacterium (EF559145)	97	Firmicutes	_	France: mesophilic digester
BS59	0.5	EU358734	Dethiobacter alkaliphilus (EF422412)	94	Firmicutes	Sulfur-reducing	Russia: soda lakes
BS60	0.5	EU358735	Clostridium sp. (AB186360)	88	Firmicutes	_	Japan: methanogenic bioreactor
BS61	0.5	EU358736	Uncultured bacterium (EF559145)	99	Firmicutes	_	France: mesophilic digester
BS62	0.5	EU358737	Uncultured bacterium (CR933151)	99	Firmicutes	_	France: naerobic sludge digester
BS63	0.5	EU358738	Alkaliflexus imshenetskii (AJ784993)	90	Bacteroidetes	Acidogenic	Russia: soda lake
BS64	0.5	EU358739	Proteiniphilum acetatigenes (AY742226)	89	Bacteroidetes	Proteolytic	China: sludge of UASB reactor
BS65	0.5	EU358740	Uncultured Bacteroidetes (AB234401)	90	Bacteroidetes	_	Japan: gut of termites
BS66	0.5	EU358741	Owenweeksia hongkongensis (AB125062)	85	Bacteroidetes	_	China: strain UST20020801
BS67	0.5	EU358742	Uncultured Bacteroidetes (AF529128)	89	Bacteroidetes	-	USA: trichloroethene-contaminated
BS68	0.5	EU358743	Bacteroidetes bacterium (AY548787)	98	Bacteroidetes	-	Finland: SRB reactor
BS69	0.5	EU358744	Alkaliflexus imshenetskii (AJ784993)	89	Bacteroidetes	Acidogenic	Russia: soda lake

OTU, operational taxonomic unit.

bacterial strains, which were Leuconostoc citreum, Clostridium quinii, Clostridium chartatabidum, Streptococcus alactolyticus, Lactobacillus reuteri and Clostridium sp. The remaining eight OTU were from uncultured bacteria. These results indicated that more than 91% bacteria in this digester were uncultured.

Within the 69 OTU, 41 were classified as Firmicutes, 16 as Bacteroides and 8 as Spirochaetes. In a phylogenetic tree (Fig. 1), 24 OTU of Firmicutes were clustered with the pure cultures belonging to Clostridia, the most of which belonged to the family Clostridiaceae. Three Firmicutes OTU were assigned to different families of Bacilli. All of the 16 Bacteroides OTU clustered together and divided into four subgroups, belonging to family Porphyromonadaceae, Rikenellaceae and two uncultured Bacteroidetes. In the eight Spirochaetes OTU, two clustered together with genus Treponema, five with genus Spirochaeta and the remaining one with uncultured Leptospiracheae. The remaining four sequences were grouped together with uncultured Fibrobacteres, Xanthomonas vasicola, uncultured planctomycete and uncultured Verrucomicrobia, respectively.

The phylogenetically most closely matched bacteria were mostly detected from the intestine of pig or other animals, waste-water treatment plant (sludge or biofilm), anaerobic reactor or digester or landfill leachate and compost, which were all related to anaerobic fermentation (Table 1). The matching micro-organisms of most OTU belonged to phylum *Firmicutes*, *Bacteroides* and *Spirochaetes*, which were fermentative acidogens. Several OTU from *Firmicutes* were cellulolytic and those mainly from *Bacteroides* were proteolytic. The most closely matched micro-organisms of three OTU from *Firmicutes* were homoacetogens and that of another one was the sulfur-reducing bacteria (Table 1).

Microbial community analysed by clone library-archaeal community

The 16S rDNA genes amplified from the total DNA extracted from biogas slurry sample with a specific archaeal primer set of 1Af/1100Ar were ligated into the pMD-18 T vector to construct a library. Exactly 186 clones were used for the following analyses and the length of the amplified genes was approximately 1·1 kbp. All of the sequences were classified into 25 OTU (Table 2). The coverage of the library was 94·1%. OTU AS01, AS02 and AS04 accounted for 27·4%, 29·0% and 10·8% of the sequenced clones in the archaeal library, respectively.

Only 6 of the 25 OTU had matches to most closely related sequences in the databases at a similarity more

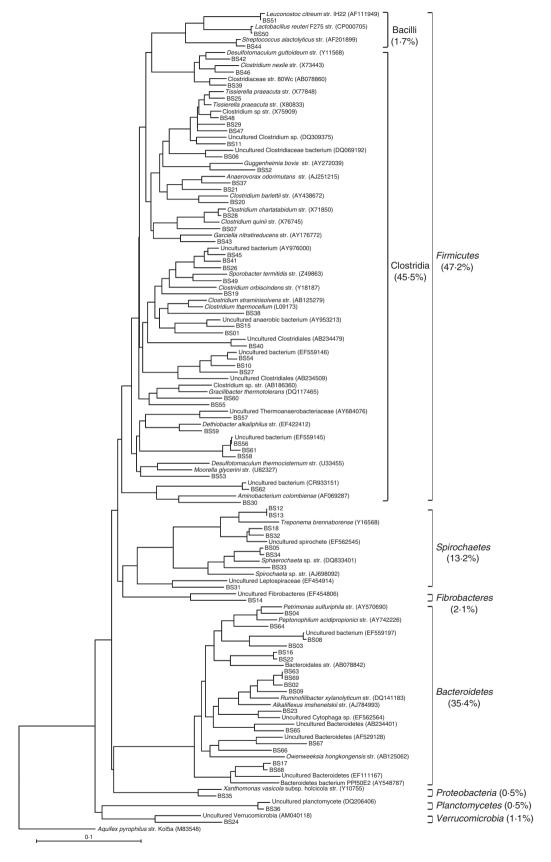


Table 2 Taxonomic relationship of archaeal 16S rDNA sequences from Chongming biogas digester compared (BLAST) with public databases (RDP, Greengenes and NCBI)

OTU	% of	Accession	Phylogenetically most closely related organism	Sm		Functional	
(%)	total	number	(accession no.)	(%)	Genus	group	Source
AS01	27-4	EU358651	Methanosarcina barkeri (AJ002476)	97	Methanosarcina	Hydrogenotrophic/ aceticlastic	New Zealand: cow
AS02	29.0	EU358652	Methanoculleus bourgensis (AB065298)	97	Methanoculleus	Hydrogenotrophic	DSM 6216
AS03	0.5	EU358653	Methanospirillum hungatei (M60880)	93	Methanospirillum	Hydrogenotrophic	_
AS04	10.8	EU358654	Methanospirillum hungatei (M60880)	97	Methanospirillum	Hydrogenotrophic	_
AS05	5.4	EU358655	Methanomicrobiales archaeon (DQ280483)	97	Methanogenium	-	USA: Skan bay
AS06	3.9	EU358656	Methanospirillum hungatei (M60880)	92	Methanospirillum	Hydrogenotrophic	_
AS07	1.1	EU358657	Methanogenium marinum (DQ177345)	95	Methanogenium	Hydrogenotrophic	USA: Skan bay
AS08	2.2	EU358658	Methanogenium marinum (DQ177344)	89	Methanogenium	Hydrogenotrophic	USA: Skan bay
AS09	0.5	EU358659	Methanothrix soehngenii (X51423)	99	Methanosaeta	Aceticlastic	Netherlands: Opfikon
AS10	0.5	EU358660	Uncultured archaeon (AY835414)	92	_	_	USA: sediment
AS11	0.5	EU358661	Methanogenium cariaci (M59130)	94	Methanogenium	Hydrogenotrophic	Library: DSM 1497
AS12	2.2	EU358662	Methanoculleus sp. (AJ550158)	94	_	Hydrogenotrophic	Germany: dm2
AS13	0.5	EU358663	Methanoculleus bourgensis (AB065298)	90	_	Hydrogenotrophic	DSM 6216
AS14	0.5	EU358664	Methanosarcina sp. HB-1 (AB288262)	91	-	Hydrogenotrophic/ aceticlastic	Japan: sedimentary rock
AS15	1.6	EU358665	Methanosarcina mazei (NC_003901)	90	-	Hydrogenotrophic/ aceticlastic	USA: strain Go1
AS16	0.5	EU358666	Methanosarcina barkeri (AF028692)	91	-	Hydrogenotrophic/ aceticlastic	France: Sar
AS17	0.5	EU358667	Methnosarcina siciliae (U89773)	96	Methanosarcina	Hydrogenotrophic/ aceticlastic	USA: C2J
AS18	1.1	EU358668	Methanoculleus bourgensis (AY196674)	93	_	Hydrogenotrophic	Australia: MS2
AS19	4.9	EU358669	Uncultured euryarchaeote (EF552190)	98	Methanosarcina	_	France: digester
AS20	0.5	EU358670	Methanospirillum hungatei (CP000254)	92	Methanospirillum	Hydrogenotrophic	USA: JF-1
AS21	1.6	EU358671	Methanospirillum sp. (AJ133792)	92	Methanospirillum	Hydrogenotrophic	Germany: Bremen
AS22	0.5	EU358672	Uncultured Methanomicrobiales (AB353214)	95	-	_	Japan: mesophilic sludge
AS23	1.1	EU358673	Methanoculleus bourgensis (AB065298)	94	_	Hydrogenotrophic	DSM 6216
AS24	0.5	EU358674	Methanoculleus bourgensis (AB065298)	95	_	Hydrogenotrophic	DSM 6216
AS25	2.2	EU358675	Methanoculleus bourgensis (AY196674)	94	_	Hydrogenotrophic	Australia: MS2

than 97% (Table 2), among which 5 OTU were matched to cultured isolates. Eighteen of the twenty-five OTU were matched to the closest related known sequences at a similarity index of between 90% and 96%; the remainders were matched to the closest related known sequences at a similarity index of 89%. More than 80·0% of the archaeal OTU in this digester may be uncultured.

All of the archaeal OTU were classified into *Methanomicrobia* of phylum *Euryarchaeota* and assigned to two branches: *Methanomicrobiales* and *Methanosarcinales* (Fig. 2). In branch *Methanomicrobiales*, four OTU were

clustered with two isolates from genus of *Methanogenium*, one OTU was grouped to genus *Methanoculleus*, five to genus *Methanospirillum* and the remaining seven were divided into four different sub-branches, which might represent new genus or new families. Within the *Methanosarcinales* branch, three OTU clustered with genus *Methanosarcina*, one OTU belonged to genus *Methanosaeta* and the remaining four (AS10, AS14, AS15 and AS16) were clustered together but separated from genus *Methanosarcina*. It suggested that OTU AS10, AS14, AS15 and AS16 might represent a new genus of *Methanosarcinaceae*.

Figure 1 Phylogenetic tree of bacteria. The tree was constructed with the neighbour-joining method of the ARB programme package using nearly complete sequences of the 16S rRNA gene. Scale bar is 10% of the estimated difference in nucleotide sequence position. *Aquifex pyrophilus* was used as the outgroup.

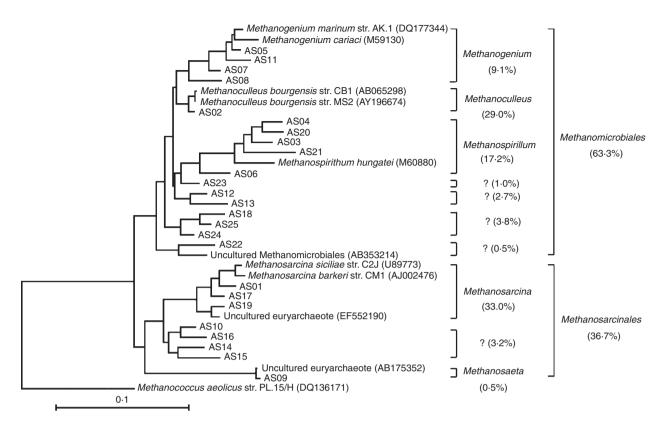


Figure 2 Phylogenetic tree of archaea. The tree was constructed using the neighbour-joining method of the ARB programme package using nearly complete sequences of 16S rRNA gene. Scale bar indicates 10% estimated difference in nucleotide sequence position. *Methanococcus aeolicus* was used as the outgroup.

However, the three most abundant OTU, AS02 (29·0%), AS01 (27·4%) and AS04 (10·8%), comprising more than 67% clones in the library, matched pure cultures of *Methanoculleus bourgensis, Methanosarcina barkeri* and *Methanospirillum hungatei* at a similarity index of 97%, respectively. This indicates that the four methanogens were the dominant archaeal species in this biogas digester.

The phylogenetically assigned archaea OTU were divided into three functional groups. Most OTU, including the most and the third most abundant OTU, were hydrogenotrophic methanogens (Table 2). The most closely matched archaea to five OTU, including the second most abundant OTU, were hydrogenotrophic/aceticlastic methanogens, and one OTU was an aceticlastic methanogen.

Structure of the dominant microbiota in the digester analysed by DGGE fingerprinting and sequencing the DGGE bands

The structures of the dominant bacteria and archaea in the slurry sample were analysed by DGGE fingerprinting (Fig. 3). Twenty-four detected bands were found in the bacterial DGGE profile (CB1–CB24) and nine bands (CA1–CA9) in the archaeal profile (Fig. 3). These results demonstrated a higher diversity of bacteria than archaea in the digester.

All of the detected bands from bacterial and archaeal DGGE fingerprinting were excised, amplified and cloned. The mobility of the inserted fragments of three clones randomly selected from each clone library (33 libraries in total) were checked via DGGE and compared with the original DGGE pattern. Clones of the inserts that migrated to the same locations as the original bands in the DGGE profile were sequenced. In total, 34 different sequences from the bacterial libraries and 11 sequences from archaeal libraries were obtained (Tables 3 and 4). Multi-fragments were found in a single band in at least 12 of the 33 recovered DGGE bands.

Nineteen of the bacterial sequences were assigned to *Firmicutes*, ten of the sequences were clustered to *Bacteroidetes*, three were grouped to *Spirochaetes* and one to *Proteobacteria* and *Verrucomicrobiales*, respectively (Table 3). Seven of the 34 bacterial sequences matched one isolate at a similarity index of $\geq 97\%$, the remainder

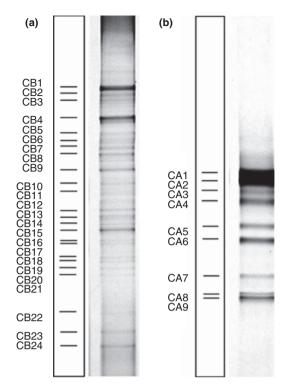


Figure 3 Denaturing gradient gel electrophoresis (DGGE) fingerprints of the bacterial (a) and archaeal (b) communities of biogas slurries obtained at Chongming in June 2005. UPGMA tree representing the genetic similarity of the microbial community profiles was obtained by polymerase chain reaction-DGGE.

matched isolates at a similarity index of <97% (Table 3), indicating that nearly 80% of the dominant bacteria in the biogas digester were unclassified.

In the DGGE profile, most of the fragments representing V3 regions of phylum *Firmicutes* were found to melt at high denaturant concentration areas, and that of phylum *Bacteroidetes* tended to melt at low denaturant concentration areas (Table 3 and Fig. 3).

Three archaeal sequences were assigned to genera *Methanoculleus* and *Methanosarcina*, respectively; one belonged to genus *Methanospirillum*, three grouped to order *Methanomicrobiales* and the remaining was grouped to order *Methanosarcinales* (Table 4). About half of the sequences were matched to uncultured taxa.

The consistency of dominant bacterial and archaeal composition revealed by clone library and DGGE profile analysis

Eighteen bacterial sequences representing 16 of the 24 recovered DGGE bands were also present in the clone library. On the other hand, seven of the nine OTU with abundance >3% were matched with the V3 sequences recovered from the DGGE bands (Table 3). This means

that 78% dominant OTU (with abundance >3%) in the bacterial clone library could be also detected by DGGE profile analysis.

Ten archaeal sequences representing all of the nine recovered DGGE bands were in the clone library. Five of the dominant six archaeal OTU with abundance more than 3% were found to match with the V3 sequences recovered from DGGE bands. Exactly 83% of the dominant OTU (with abundance >3%) in the archaeal library were observed in the DGGE profile.

Discussion

Cloning and sequencing the full length of 16S rDNA has been frequently applied to elucidate the exact composition of a microbial community. However, DGGE profiles of the V3 regions can be employed to reflect the dominant community structure and to analyse the structural variation between different systems or the dynamics of the same system (Muyzer et al. 1993). In this study, sequencing of V3 regions from recovered DGGE bands and 16S rDNA from cloning libraries were both used to analyse the microbial community in an anaerobic digester. Both analyses indicated that the community was mainly composed of phyla Firmicutes, Bacteroides and Spirochaetes for bacteria, and orders Methanomicrobiales and Methanosarcinales for archaea. The analyses also indicated that the diversity of bacteria was higher than that of achaera. Importantly, 78% of dominant OTU (with abundance >3%) in the bacterial library and 83% of dominant OTU in the archaeal library could be detected in the DGGE profile. These results indicate that the DGGE profile in this study clearly reflected the dominant composition of the microbial community in the Qianwei biogas digester. Additionally, the major components of this bacterial community could be separately located on the different areas in the DGGE profile, e.g. Firmicutes tended to appear at the high denaturant concentration area and Bacteroidetes at low denaturant concentration areas in the DGGE profile. The analysis of DGGE profiles may be useful in displaying the dominant microbial composition in biogas digesters in our further studies. However, more OTU were identified by sequencing 16S rDNA libraries, indicating that DGGE may underperform in elucidating the diversity or the exact composition of a complex microbial community.

Anaerobic digesters are widely used to treat different wastes, e.g. brewery and pulp industry wastewater containing different carbohydrates, long-chain fatty acids, volatile fatty acids, methanethiol, terephthalate (Mata-Alvarez et al. 2000; Yadvika et al. 2004). Many studies focus on analyses of microbial community in such anaerobic digesters by using different approaches (Cotta et al. 2003; Snell-Castro et al. 2005; Peu et al. 2006). In these studies, the

Table 3 Bacterial 16S rDNA sequence similarities of excised bands that appear in Fig. 3

		1					
		Phylogenetically					
		most closely					
	Accession	related organism	Lg	Sm1	Corresponding	Sm2	Functional
Band ID	number	(accession no.)	(bp)	(%)	OTU no.	(%)	group
CB1	EU358617	Proteiniphilum acetatigenes (AY742226)	189	95	BS04, BS64	100	Proteolytic
CB2	EU358618	Eubacterium tortuosum (L34683)	195	83	_	_	_
CB3	EU358619	Petrimonas sulfuriphila (AY570690)	189	89	BS03	99	Acidogenic
CB4-1	EU358620	Uncultured bacterium (EF559197)	189	100	BS08	100	_
CB4-2	EU358621	Alkaliflexus imshenetskii (AJ784993)	160	89	BS02, 63, 69	100	Acidogenic
CB5-1	EU358622	Uncultured bacterium (EF686929)	189	97	BS08	98	_
CB5-2	EU358623	Uncultured bacterium (AY816908)	190	95	_	_	_
CB6-1	EU358624	Spirochaeta sp. grapes (AF357917)	194	87	_	_	_
CB6-2	EU358625	Uncultured bacterium (AJ937700)	189	95	_	_	_
CB7	EU358626	Uncultured bacterium (AB290394)	189	99	_	_	_
CB8	EU358627	Uncultured bacterium (AJ628010)	189	94	BS23	100	_
CB9-1	EU358628	Uncultured spirochete (EF562545)	194	96	BS32	99	_
CB9-2	EU358629	Clostridium orbiscindens(Y18187)	174	95	_	_	Acidogenic
CB10-1	EU358630	Syntrophomonas zehnderi (DQ898277)	194	93	_	_	Acetogenic
CB10-2	EU358631	Oscillospira guilliermondii (AB040499)	171	100	_	_	_
CB11-1	EU358632	Clostridium intestinale (X76740)	168	96	_	_	_
CB11-2	EU358633	Uncultured bacterium (AY816908)	190	97	BS63, 69	98	_
CB12	EU358634	Acetivibrio cellulolyticus (L35516)	169	99	_	_	Cellulolytic
CB13-1	EU358635	Uncultured bacterium (EF559146)	171	98	BS54	100	_
CB13-2	EU358636	Clostridium thermocellum (L09173)	169	92	_	_	Cellulolytic
CB14	EU358637	Succinivibrio dextrinosolvens (Y17600)	171	96	_	_	Acidogenic
CB15	EU358638	Uncultured Clostridiaceae (DQ069192)	169	100	BS06	100	_
CB16-1	EU358639	Clostridium hydroxybenzoicum (L11305)	169	94	_	_	_
CB16-2	EU358640	Tissierella praeacuta (X80833)	169	96	BS47	99	_
CB17	EU358641	Clostridium chartatabidum (X71850)	169	98	BS28	98	Cellulolytic
CB18	EU358642	Uncultured Verrucomicrobiales (AJ853598)	194	96			_
CB19-1	EU358643	Anaerovorax odorimutans (AJ251215)	172	98	BS21	99	Acidogenic
CB19-2	EU358644	Uncultured bacterium (AY980698)	172	96	_	_	_
CB20	EU358645	Clostridium quinii (X76745)	169	100	BS07	100	Acidogenic
CB21-1	EU358646	Clostridium quinii (X76745)	169	98	BS07	98	Acidogenic
CB21-2	EU358647	Tissierella praeacuta (X80833)	169	95	BS29	98	_
CB22	EU358648	Uncultured bacterium (EF559145)	195	95	BS58	99	_
CB23	EU358649	Spirochaeta sp. SPN1 (AJ698092)	194	95	BS05	99	_
CB24	EU358650	Dethiosulfovibrio acidaminovorans (AY005466)	172	97	_	_	Sulfur-reducing

Sm1, similarity between the sequences of denaturing gradient gel electrophoresis (DGGE) band and its phylogenetically closely related organism. Sm2, similarity between the sequences of DGGE band and its corresponding operational taxonomic unit (OTU).

microbial structures in these systems varied greatly owing to the difference of the substrates used. Firmicutes, Nitrospira and Deferribacteres were found to be the predominant bacteria and Methanosaeta concilii was the dominant methanogenic archaea in an anaerobic digester treating wastewater from a beer brewery (Diaz et al. 2006). In an anaerobic digester treating long-chain fatty acids, species of Syntrophomonadaceae and Syntrophobacteraceae families, which oxidize fatty acids, were the predominant bacteria (Sousa et al. 2007b). When degrading methanethiol from paper mill wastewater, methylotrophic methanogens Methanomethylovorans hollandica were enriched (de Bok et al. 2006). In a laboratory methanogenic digester amended with glucose, Spirochaetes-, eubacterium- and

propionibacterium-like bacteria were found to be dominant (Fernandez *et al.* 1999, 2000). Pig manure, which was mainly comprised of undigested biomass and some fatty acids, such as acetic acid and propionic acid, was the sole substrate in Qianwei biogas digester analysed in this study. *Firmicutes* (47·2%), *Bacteroides* (35·4%) and *Spirochaetes* (13·2%) were found to be the three most abundant bacterial phyla in this study. Within phylum *Firmicutes*, class *Clostridia* was the most dominant of the bacterial community (45·5% of the clones).

Several other studies have shown that within the bacterial and archaeal community of a pig manure slurry and a manure storage pit, *Eubacterium*, *Clostidium*, *Bacillus–Lactobacillus–Streptococcus* subdivision, *Mycoplasma* and

Table 4 Archaeal 16S rDNA sequence similarities of excised bands that appear in Fig. 3

Band ID	Accession number	Phylogenetically most closely related organism (accession no.)	Lg (bp)	Sm1 (%)	Corresponding OTU no.	Sm2 (%)	Functional group
CA1-1	EU358606	Methnosarcina siciliae (U89773)	152	98	AS17	100	Hydrogenotrophic/ aceticlastic
CA1-2	EU358607	Methanosarcina barkeri (AJ002476)	152	97	AS01	98	Hydrogenotrophic/ aceticlastic
CA2-1	EU358608	Methanoculleus bourgensis (AB065298)	148	97	AS02, 18, 24, 25	99	Hydrogenotrophic
CA2-2	EU358609	Uncultured archaeon (AM712547)	132	98	_	_	_
CA3	EU358610	Uncultured Methanomicrobiales (AB353214)	148	96	AS22	99	_
CA4	EU358611	Uncultured Methanomicrobiales (AB353214)	148	97	AS22	99	_
CA5	EU358612	Methanoculleus bourgensis (AB065298)	148	98	AS13	99	Hydrogenotrophic
CA6	EU358613	Methanoculleus bourgensis (AB065298)	148	98	AS13	99	Hydrogenotrophic
CA7	EU358614	Methanospirillum sp. (AJ133792)	148	95	AS3, 4, 6, 15, 20, 21	99	Hydrogenotrophic
CA8	EU358615	Uncultured Methanosarcinales (AB353215)	152	99	AS17	99	Hydrogenotrophic/ aceticlastic
CA9	EU358616	Uncultured euryarchaeote (EF552190)	152	99	AS19	99	_

Sm1, similarity between the sequences of denaturing gradient gel electrophoresis (DGGE) band and its phylogenetically closely related organism. Sm2, similarity between the sequences of DGGE band and its corresponding operational taxonomic unit (OTU).

the Flexibacter–Cytophaga–Bacteroides were the main components of the bacterial communities (Snell-Castro et al. 2005; Peu et al. 2006). Hydrogenotrophic methanogens, such as Methanoculleus, Methanogenium and Methanobrevibacter, dominated the archaeal communities (Whitehead and Cotta 1999; Tang et al. 2004; Hori et al. 2006). This agrees with the results from our digester, suggesting most archaea in our digester might originate from pig manure.

Most phylogenetically closely matched bacteria to the OTU identified in the biogas digester were found in anaerobic environments, such as the guts of animal or insects, sediments, anaerobic digesters and faeces (Table 1). Many phylogenetically closest matched taxa were uncultured organisms or function-unidentified organisms. The metabolic functions of their related OTU in the biogas digester were unknown. Among the function-identified bacteria, most were acidogenic, producing H₂, CO₂, formate, acetate and other fatty acids as well as a small amount of ethanol from cellobiose, D-fructose, N-acetylglucosamine, D-glucose, maltose, mannose and saccharose (Vandamme et al. 1999); several OTU related to the Clostridium sp. might be cellulolytic and the other three ones were related to proteolytic, both of which might take charge of decomposing polymers in the pig manure to monomers; Only three OTU might be homoacetogenic. This seems to be consistent with the fact that most archaeal OTU or the most abundant OTU were hydrogenotrophic methanogens, and only one OTU at low frequency was identified as the aceticlastic methanogen (Table 2).

The phylum Euryarchaeota was the major methanogenic archaeal group in anaerobic fermentation environments. Methanoculleus bourgensis, Methanosarcina barkeri, Methanospirillum hungatei and Methanomicrobiales archaeon were the most abundant methanogenic species in our digester (Table 2). Each of them showed some specific characteristics in methanogenic metabolism. Methanoculleus bourgensis was reported to use H2-CO2 or formate as a substrate for growth and methanogenesis, and is a hydrogenotrophic methanogen (Blotevogel et al. 1992). Methanospirillum hungatei produces methane only from H₂-CO₂ or formate, but not from acetate or ethanol and methanol, being a strictly hydrogenotrophic methanogen (Ferry et al. 1974). Methanosarcina barkeri could be used in different substrates to produce methane, including H₂-CO₂, methanol, mono-, di- and trimethylamines, acetate and CO (Bryant and Boone 1987), and is a hydrogenotrophic or aceticlastic methanogen.

Chimeric sequences of the full length of 16S rDNA were usually found while analysing the compositions of complex microbial communities (Ashelford *et al.* 2006). Therefore, several programmes were designed to identify chimeric sequences (Cole *et al.* 2003; DeSantis *et al.* 2006). Ashelford *et al.* (2006) reported that the average chance to falsely identify a sequence as chimeric by the Bellerophon programme was 7·2%. Fifty-three 16S rDNA sequences from bacteria were identified to be putatively chimeric. To ensure most of them to be assigned rightly, the putative chimeric sequences were confirmed by two different programmes, and some of them were further confirmed by PCR amplification. However, it was still

possible that some of the 53 sequences be assigned as chimeric by wrong, and thus might underestimate the bacterial diversity in the biogas digester.

High diversities of microbial composition and metabolism were found in this microbiota in this study. It provides a pool of functional micro-organisms involved in biomass transformation. However, most micro-organisms, more than 91% bacteria and 80% of archaea, in this digester were uncultured. Presently it is impossible to elucidate the total metabolic process in the biogas digester using only the analysis of microbial composition. The recent application of metagenomic techniques suggest they may provide a new approach to obtain function genes related to biomass transformation within the assemblage. A lab anaerobic fermentation system and its control system are necessary in elucidating the relationship between the biogas-producing efficiency and microbiota composition under variable substrate conditions, with the addition of an inhibitor or accelerant, and/or adding specific micro-organisms for bio-augmentation.

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References

- Angenent, L.T., Karim, K., Al-Dahhan, M.H., Wrenn, B.A. and Domíguez-Espinosa, R. (2004) Production of bioenergy and biochemicals from industrial and agricultural wastewater. *Trends Biotechnol* **22**, 477–485.
- Ashelford, K.E., Chuzhanova, N.A., Fry, J.C., Jones, A.J. and Weightman, A.J. (2006) New screening software shows that most recent large 16S rRNA gene clone libraries contain chimeras. *Appl Environ Microbiol* 72, 5734–5741.
- Bianciotto, V., Bandi, C., Minerdi, D., Sironi, M., Tichy, H.V. and Bonfante, P. (1996) An obligately endosymbiotic fungus itself harbors obligately intracellular bacteria. *Appl Environ Microbiol* 62, 3005–3010.

- Blotevogel, K.H., Gahl-Janßen, R., Jannsen, S., Fischer, U., Pilz, F., Auling, G., Macario, A.J.L. and Tindall, B.J. (1992) Isolation and characterization of a novel mesophilic, freshwater methanogen from river sediment *Methanoculleus oldenburgensis* sp. nov. *Arch Microbiol* 157, 54–59.
- de Bok, F.A.M., van Leerdam, R.C., Lomans, B.P., Smidt, H., Lens, P.N.L., Janssen, A.J.H. and Stams, A.J.M. (2006) Degradation of methanethiol by methylotrophic methanogenic archaea in a lab-scale upflow anaerobic sludge blanket reactor. *Appl Environ Microbiol* 72, 7540–7547.
- Boone, D.R. and Bryant, M.P. (1980) Propionate-degrading bacterium, *Syntrophobacter wolinii* sp. nov., gen. nov., from methanogenic ecosystems. *Appl Environ Microbiol* **40**, 626–632.
- Bryant, M.P. and Boone, D.R. (1987) Emended description of strain MS^T (DSM 800^T), the type strain of *Methanosarcina* barkeri. Int J Syst Bacteriol 37, 169–170.
- Chynoweth, D.P., Wilkie, A.C. and Owens, J.M. (1999) Anaerobic treatment of piggery slurry-review. *Asian–Austral I Anim Sci* 12, 607–628.
- Cirne, D., Lehtomäki, A., Björnsson, L. and Blackall, L.L. (2007) Hydrolysis and microbial community analysis in two-stage anaerobic digestion of energy crops. J Appl Microbiol 103, 516–527.
- Cole, J.R., Chai, B., Marsh, T.L., Farris, R.J., Wang, Q., Kulam, S.A., Chandra, S., McGarrell, D.M. *et al.* (2003) The ribosomal database project (RDP-II): previewing a new autoaligner that allows regular updates and the new prokaryotic taxonomy. *Nucleic Acids Res* **31**, 442–443.
- Connaughton, S., Collins, G. and O'Flaherty, V. (2006) Development of microbial community structure and activity in a high-rate anaerobic bioreactor at 18°C. *Water Res* **40**, 1009–1017.
- Cotta, M.A., Whitehead, T.R. and Zeltwanger, R.L. (2003) Isolation, characterization and comparison of bacteria from swine faeces and manure storage pits. *Environ Microbiol* **5**, 737–745.
- DeSantis, T.Z., Hugenholtz, P., Larsen, N., Rojas, M., Brodie, E.L., Keller, K., Huber, T., Alevi, D. *et al.* (2006) Greengenes, a chimera-checked 16S rRNA gene database and workbench compatible with ARB. *Appl Environ Microbiol* 72, 5069–5072.
- Diaz, E.E., Stams, A.J.M., Amils, R. and Sanz, J.L. (2006) Phenotypic properties and microbial diversity of methanogenic granules from a full scale upflow anaerobic sludge bed reactor from a full-scale anaerobic sludge bed reactor treating brewery wastewater. Appl Environ Microbiol 72, 4942–4949.
- Drake, H.L., Küsel, K. and Matthies, C. (2006) Acetogenic prokaryotes. In *The Prokaryotes*, 3rd edn ed. Dworkin, M., Falkow, S., Rosenberg, E., Schleifer, K.-H. and Stackebrandt, E. pp. 354–420 New York, NY: Springer-Verlag.
- Embley, T.M., Finlay, B.J., Thomas, R.H. and Dyal, P.L. (1992) The use of rRNA sequences and fluorescent probes to investigate the phylogenetic positions of the anaerobic

- ciliate *Metopus palaeformis* and its archeabacterial endosymbiont. *J Gen Microbiol* **138**, 1479–1487.
- Fernandez, A., Huang, S., Seston, S., Xing, J., Hickey, R., Criddle, C. and Tiedje, J. (1999) How stable is stable? Function versus community composition *Appl Environ Microbiol*. 65, 3697–3704.
- Fernandez, A.S., Hashsham, S.A., Dollhopf, S.L., Raskin, L., Glagoleva, O., Dazzo, F.B., Hickey, R.F., Criddle, C.S. et al. (2000) Flexible community structure correlates with stable community function in methanogenic bioreactor communities perturbed by glucose. Appl Environ Microbiol 66, 4058–4067.
- Ferry, J.G., Smith, P.H. and Wolfe, R.S. (1974) *Methanospirillum*, a new genus of methanogenic bacteria, and characterization of *Methanospirillum hungatii* sp. nov. *Int J Syst Bacteriol* **24**, 465–469.
- Freeman, S.A., Sierra-Alvarez, R., Altinbas, M., Hollingsworth, J., Stams, A.J. and Smidt, H. (2008) Molecular characterization of mesophilic and thermophilic sulfate reducing microbial communities in expanded granular sludge bed (EGSB) reactors. *Biodegradation* 19, 161–177.
- Gupta, C.L. (2003) Role of renewable energy technologies in generating sustainable livelihoods. *Renew Sust Energ Rev* 7, 155–174.
- Hao, Y.Y. and Liu, R.H. (2006) Test and analysis of the raw material of biogas fermentation and its by-product. *J Anhui Agricult Sci* 34, 3429–3431.
- Hori, T., Haruta, S., Ueno, Y., Ishii, M. and Igarashi, Y. (2006) Dynamic transition of a methanogenic population in response to the concentration of volatile fatty acids in a thermophilic anaerobic digester. *Appl Environ Microbiol* 72, 1623–1630.
- Karakashev, D., Batstone, D.J. and Angelidaki, I. (2005) Influence of environmental conditions on methanogenic compositions in anaerobic biogas reactors. *Appl Environ Microbiol* 71, 331–338.
- Klocke, M., Mähnert, P., Mundt, K., Souidi, K. and Linke, B. (2007) Microbial community analysis of a biogas-producing completely stirred tank reactor fed continuously with fodder beet silage as mono-substrate. Syst Appl Microbiol 30, 139–151.
- Krave, A.S., Lin, B., Braster, M., Laverman, A.M., van Straalen, N.M., Röling, W.F. and van Verseveld, H.W. (2002) Stratification and seasonal stability of diverse bacterial communities in a *Pinus merkusii* (pine) forest soil in central Java, Indonesia. *Environ Microbiol* 4, 361–373.
- Ley, R.E., Backhed, F., Turnbaugh, P., Lozupone, C.A., Knight, R.D. and Gordon, J.I. (2005) Obesity alters gut microbial ecology. *Proc Natl Acad Sci USA* 102, 11070–11075.
- Ludwig, W., Strunk, O., Westram, R., Richter, L., Meier, H., Yadhukumar, A., Buchner, A., Lai, T. et al. (2004) ARB: a software environment for sequence data. *Nucleic Acids Res* 32, 1363–1371.
- Mata-Alvarez, J., Mace, S. and Llabres, P. (2000) Anaerobic digestion of organic solid wastes. An overview of research

- achievements and perspectives. *Bioresour Technol.* **74**, 3–16.
- Meher, K.K. and Ranade, D.R. (1993) Isolation of propionate degrading bacterium in co-culture with a methanogen from a cattle dung biogas plant. *J Biosci* **18**, 271–277.
- Mladenovska, Z., Dabrowski, S. and Ahring, B.K. (2003)
 Anaerobic digestion of manure and mixture of manure with lipids: biogas reactor performance and microbial community analysis. *Water Sci Technol* **48**, 271–278.
- Mladenovska, Z., Hartmann, H., Kvist, T., Sales-Cruz, M., Gani, R. and Ahring, B.K. (2006) Thermal pretreatment of the solid fraction of manure: impact on the biogas reactor performance and microbial community. *Water Sci Technol* **53**, 59–67.
- Muyzer, G., de Waal, E.C. and Uitterlinden, A.G. (1993)
 Profiling of complex microbial populations by denaturing gradient gel electrophoresis analysis of polymerase chain reaction-amplified genes encoding for 16S rRNA. *Appl Environ Microbiol* **59**, 695–700.
- Ney, U., Macario, A.J., Conway, L., DE Macario, E., Aivasidis, A., Schoberth, S.M. and Sahm, H. (1990) Quantitative microbiological analysis of bacterial community shifts in a high-rate anaerobic bioreactor treating sulfite evaporator condensate. *Appl Environ Microbiol* 56, 2389–2398.
- Ohkuma, M. and Kudo, T. (1996) Phylogenetic diversity of the intestinal bacterial community in the termite *Reticulitermes speratus*. *Appl Environ Microbiol* **62**, 461–468.
- Øvreås, L., Forney, L., Daae, F.L. and Torsvik, V. (1997) Distribution of bacterioplankton in meromictic Lake Saelenvannet, as determined by denaturing gradient gel electrophoresis of PCR-amplified gene fragments coding for 16S rRNA. Appl Environ Microbiol 63, 3367–3373.
- Peu, P., Brugere, H., Pourcher, A.M., Kerouredan, M., Godon, J.J., Delgenes, J.P. and Dabert, P. (2006) Dynamics of a pig slurry microbial community during anaerobic storage and management. *Appl Environ Microbiol* 72, 3578–3585.
- Pretty, J.N., Ball, A.S., Li, X.Y. and Ravindranath, N.H. (2002) The role of sustainable agriculture and renewable-resource management in reducing greenhouse-gas emissions and increasing sinks in China and India. *Phil Trans R Soc Lond A* **360**, 1741–1761.
- Schloss, P.D. and Handelsman, J. (2005) Introducing DOTUR, a computer program for defining operational taxonomic units and estimating species richness. *Appl Environ Micro-biol* 71, 1501–1506.
- Snell-Castro, R., Godon, J.J., Delgenes, J.P. and Dabert, P. (2005) Characterisation of the microbial diversity in a pig manure storage pit using small subunit rDNA sequence analysis. FEMS Microbiol Ecol 52, 229–242.
- Sousa, D.Z., Pereira, M.A., Smidt, H., Stams, A.J.M. and Alves, M.M. (2007a) Molecular assessment of complex microbial communities degrading long chain fatty acids (LCFA) in methanogenic bioreactors. *FEMS Microbiol Ecol* **60**, 252–265.

- Sousa, D.Z., Pereira, M.A., Stams, A.J.M., Alves, M.M. and Smidt, H. (2007b) Microbial communities involved in anaerobic degradation of unsaturated or saturated long-chain fatty acids. *Appl Environ Microbiol* **73**, 1054–1064.
- Tang, Y., Shigematsu, T., Ikbal, Morimura, S. and Kida, K. (2004) The effects of micro-aeration on the phylogenetic diversity of microorganisms in a thermophilic anaerobic municipal solid-waste digester. Water Res 38, 2537–2550.
- Thompson, J.R., Marcelino, L.A. and Polz, M.F. (2002) Heteroduplexes in mixed-template amplifications: formation, consequence and elimination by 'reconditioning PCR'. *Nucleic Acids Res* **30**, 2083–2088.
- Vandamme, P., Devriese, L.A., Haesebrouck, F. and Kersters, K. (1999) *Streptococcus intestinalis* (Robinson *et al.* 1988) and *Streptococcus alactolyticus* (Farrow *et al.* 1984) are phenotypically indistinguishable. *Int J Syst Bacteriol* **49**, 737–741.
- Wakase, S., Sasaki, H., Itoh, K., Otawa, K., Kitazume, O., Nonaka, J., Satoh, M., Sasaki, T. et al. (2008) Investigation of the microbial community in a microbiological additive used in a manure composting process. *Bioresour Technol* 99, 2687–2693.

- Whitehead, T.R. and Cotta, M.A. (1999) Phylogenetic diversity of methanogenic archaea in swine waste storage pits. *FEMS Microbiol Lett* **179**, 223–226.
- Yadvika, S., Sreekrishnan, T.R., Kohli, S. and Rana, V. (2004) Enhancement of biogas production from solid substrates using different techniques – a review. *Bioresour Technol* **95**, 1–10.
- Ye, N.F., Lü, F., Shao, L.M., Godon, J.J. and He, P.J. (2007) Bacterial community dynamics and product distribution during pH-adjusted fermentation of vegetable wastes. *J Appl Microbiol* **103**, 1055–1065.
- Zhang, X.L., Yan, X., Gao, P.P., Wang, L.H., Zhou, Z.H. and Zhao, L.P. (2005) Optimized sequence retrieval from single bands of TGGE (temperature gradient gel electrophoresis) profiles of the amplified 16S rDNA fragments from an activated sludge system. *J Microbiol Methods* **60**, 1–11.
- Zhao, Y.Z., Zhang, H., Boone, D.R. and Mah, R.A. (1986) Isolation and characterization of a fast-growing, thermophilic methanobacterium species. *Appl Environ Microbiol* **52**, 1227–1229.